DRAFT

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Standard Test Method for Elemental, Oxidized, Particle-Bound, and Total Mercury in Flue Gas Generated from Coal-Fired Stationary Sources (Ontario Hydro Method)¹

1. Scope

1.1 This method applies to the determination of elemental, oxidized, particle-bound, and total mercury emissions from coal-fired stationary sources.

1.2 This method is applicable to elemental, oxidized, particle-bound, and total mercury concentrations ranging from approximately 0.5 to $100 \mu g/dscm$.

1.3 This method describes equipment and procedures for obtaining samples from effluent ducts and stacks, equipment and procedures for laboratory analysis, and procedures for calculating results.

1.4 This method is applicable for sampling elemental, oxidized, and particle-bound mercury at the inlet and outlet of emission control devices and for calculating control device mercury collection efficiency.

1.5 Method applicability is limited to flue gas stream temperatures within the thermal stability range of the sampling probe and filter components.

1.6 The values stated in SI units are to be regarded as the standard. The values in parentheses are for information only.

1.7 This standard does not purport to address all of the safety concerns, if any, associated with its use. It is the responsibility of the user of this standard to establish appropriate safety and health practices and determine the applicability of regulatory limitations prior to use.

1.8 This standard assumes users are familiar with EPA stack-gas sampling procedures as stated in EPA Methods 1–4, Method 5, and Method 17.

2. Referenced Documents

2.1 ASTM Standards:D 1193 Specification for Reagent Water²

¹ This test method is under the jurisdiction of ASTM Committee D-22 on Sampling and Analysis of Atmospheres and is the direct responsibility of Subcommittee D22.03 on Ambient Atmospheres and Source Emissions.

² Annual Book of ASTM Standards, Vol. 11.01.

D1356 Definitions of Terms Relating to Atmospheric Sampling and Analysis³

D 2986 Evaluation of Air-Assay Media by the Monodisperse DOP (Dioctyl Phthalate) Smoke Test³

D 3154 Test Method for Average Velocity in a Duct (Pitot Tube Method)³

D 3685 Particulates Independently or for Particulates and Collected Residue Simultaneously in Stack Gases³

E 1 Specification for ASTM Thermometers⁴

2.2 Other Standards:⁵

EPA Method 1 – Sample and Velocity Traverses for Stationary Sources

EPA Method 2 – Determination of Stack Gas Velocity and Volumetric Flow Rate (Type S Pitot Tube)

EPA Method 3 – Gas Analysis for the Determination of Dry Molecular Weight

EPA Method 4 - Determination of Moisture Content in Stack Gases

EPA Method 5 – Determination of Particulate Emissions from Stationary Sources

EPA Method 12 - Determination of Inorganic Lead Emissions from Stationary Sources

EPA Method 17– Determination of Particulate Emissions from Stationary Sources (In-Stack Filtration Method)

EPA Method 29 – Determination of Metals Emissions from Stationary Sources EPA Method 101A – Determination of Particle-Bound and Gaseous Mercury Emissions from Sewage Sludge Incinerators

EPA Method 301 – Field Validation of Pollutant Measurement Methods from Various Waste Media

EPA SW 846 7470 – Mercury in Liquid Waste – Manual Cold Vapor Technique EPA Water and Waste 600/4-79-020 – Methods for Chemical Analysis of Water and Wastes

3. Terminology

3.1 Definitions other than those given below in Sections 3.2, 3.3, and 3.4 are listed in ASTM D 1356.

3.2 *Definitions of Terms specific to the standard:*

3.2.1 *elemental mercury*—mercury in its zero oxidation state, Hg⁰.

³ Annual Book of ASTM Standards, Vol. 11.03.

⁴ Annual Book of ASTM Standards, Vol 14.02.

⁵ Available from the U.S. Environmental Protection Agency's Emission Measurement Technical Information Center or Code of Federal Regulations (40 CFR Part 60, Appendix A or 40 CFR Part 61, Appendix B).

3.2.2 *oxidized mercury*—mercury in its mercurous or mercuric oxidation states: Hg_2^{2+} and Hg^{2+} , respectively.

3.2.3 *elemental mercury catch*—mercury collected in the acidified hydrogen peroxide $(HNO_3-H_2O_2)$ and potassium permanganate $(H_2SO_4-KMnO_4)$ impinger solutions employed in this method. This is gaseous Hg⁰.

3.2.4 *oxidized mercury* catch—mercury collected in the aqueous potassium chloride (KCl) impinger solution employed in this method. This is gaseous Hg^{2+} .

3.2.5 *particle-bound mercury catch*—mercury associated with the particulate matter collected in the front half of the sampling train.

3.2.6 *sample train*—complete setup including nozzle, probe, probe liner, filter, filter holder, impingers, and connectors.

3.2.7 *Impinger train*—setup includes only the impingers and connectors.

3.2.8 *front half of the sampling train*—all mercury collected on and upstream of the sample filter.

3.2.9 *total mercury*— all mercury (solid-bound, liquid, or gaseous) however generated or entrained in the flue gas stream (i.e., summation of elemental, oxidized, and particle-bound mercury).

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22	Symbole
3.3	Symbols:

А	=	cross-sectional area of stack, $m^2(ft^2)$
B _{ws}	=	water vapor in the gas stream, proportion by volume
ΔH	=	average pressure differential across the orifice meter, kPa (in. H_2O)
Hg _{ash}	=	concentration of mercury in sample filter ash, $\mu g/g$
Hg ^{tp}	=	concentration of particle-bound mercury, µg/dscm
Hg^0	=	concentration of elemental mercury, µg/dscm
Hg^{2+}	=	concentration of oxidized mercury, µg/dscm
IR	=	instrument reading from mercury analyzer, µg/L
L _p	=	leakage rate observed during the posttest leak check, m ³ /min (cfm)
L	=	maximum acceptable leakage rate
M _s	=	molecular weight of stack gas, wet basis, g/g-mole (lb/lb-mole)
$M_{\rm w}$	=	molecular weight of water, 18.0 g/g-mole (18.0 lb/lb-mole)
Ν	=	Normal conditions, defined as 0°C and 1 atmosphere pressure (in the U.S. N and
		standard conditions are the same in SI units)
$\mathbf{P}_{\mathrm{bar}}$	=	barometric pressure at the sampling site, kPa (in. Hg)
P _s	=	absolute stack gas pressure, kPa (in. Hg)
P _{std}	=	standard absolute pressure, 101.3 kPa (29.92 in. Hg)
R		ideal gas constant, 0.008314 kPa-m ³ /K-g-mole (21.85 in. Hg-ft ³ /°R-lb-mole)
T _m	=	absolute average dry gas meter temperature, K (°R)

- T_s = absolute stack temperature, K (°R)
- T_{std} = standard absolute temperature, 293 K (528°R)
- V_D = total digested volume, mL
- V_{m} = volume of gas sample as measured by dry gas meter, dcm (dscf)
- $V_{m(std)}$ = volume of gas sample measured by the dry gas meter, corrected to standard conditions, dscm (dscf)
- $V_{w(std)}$ = volume of water vapor in the gas sample, corrected to standard conditions, scm (scf)
- W_{ash} = total mass of ash on sample filter, g
- W_{lc} = total mass of liquid collected in impingers and silica gel, g (lb)
- Y = dry gas meter calibration factor
- θ = total sampling time, min
- θ_1 = sampling time interval, from the beginning of a run until the first component change, min

4. Summary of Test Method

4.1 A sample is withdrawn from the flue gas stream isokinetically through a probe/filter system, maintained at 120°C or the flue gas temperature, whichever is greater, followed by a series of impingers in an ice bath. Particle-bound mercury is collected in the front half of the sampling train. Oxidized mercury is collected in impingers containing a chilled aqueous potassium chloride solution. Elemental mercury is collected in subsequent impingers (one impinger containing a chilled aqueous acidic solution of hydrogen peroxide and three impingers containing chilled aqueous acidic solutions of potassium permanganate). Samples are recovered, digested, and then analyzed for mercury using cold-vapor atomic absorption (CVAAS) or fluorescence spectroscopy (CVAFS).

5. Significance and Use

5.1 The measurement of particle-bound, oxidized, elemental, and total mercury in stationary-source flue gases provides data that can be used for dispersion modeling, deposition evaluation, human health and environmental impact assessments, emission reporting, compliance determinations, etc. Particle-bound, oxidized, and elemental mercury measurements before and after control devices may be necessary for optimizing and evaluating the mercury removal efficiency of emission control technologies.

6. Interferences

There are no known interferences, but certain biases may be encountered (See Section 16).

7. Apparatus

7.1 *Sampling Train*—similar to ASTM D 3685, EPA Method 5/EPA Method 17 and EPA Method 29 trains, as illustrated in Fig. 1.

7.1.1 *Probe Nozzle (Probe Tip)*—Glass nozzles are required unless alternate nozzles are constructed of materials that are free from contamination and will not interact with the sample. Probe fittings constructed of polytetrafluoroethylene (PTFE), polypropylene, etc., are required instead of metal fittings to prevent contamination.

7.1.2. *Probe Liner*—If the sample train is to be in EPA Method 5 configuration (out-of-stack filtration), the probe liner must be constructed of quartz or borosilicate glass. If an EPA Method 17 (in-stack filtration) sampling configuration is used, the probe/probe liner may be constructed of borosilicate glass, quartz or, depending on the flue gas temperature, PTFE.

7.1.3 *Pitot Tube*—Type S pitot tube. Refer to Section 2.2 of EPA Method 2 for a description.

7.1.4 *Differential Pressure Gauges*—inclined manometers or equivalent devices. Refer to Section 2.1 of EPA Method 2 for a description.

7.1.5 *Filter Holder* — constructed of borosilicate glass or PTFE-coated stainless steel with a PTFE filter support or other nonmetallic, noncontaminating support. Do not use a glass frit or stainless steel wire screen. A silicone rubber or PTFE gasket, designed to provide a positive seal against leakage from outside or around the filter, may be used.

7.1.6 *Connecting Umbilical Tube*—heated PTFE tubing. This tube must be heated to a minimum of 120°C to help prevent water and acid condensation. (The umbilical tube is defined as any tubing longer than 0.5 m that connects the filter holder to the impinger train).

7.1.7 Probe and Filter Heating System

7.1.7.1 *EPA Method 5 Configuration*—For EPA Method 5 configuration, the temperature of the flue gas, sample probe, and the exit of the sample filter must be monitored using temperature sensors capable of measuring temperature to within $3^{\circ}C$ ($5.4^{\circ}F$). The heating system must be capable of maintaining the sample gas temperature of the probe and exit of the sample filter to within $\pm 15^{\circ}C$ ($\pm 27^{\circ}F$) of the flue gas temperature. Regardless of the flue gas temperature, to prevent water and acid condensation, at no time must the probe temperature, sample filter exit gas temperature, or the temperature of the connecting umbilical cord be less than $120^{\circ}C$.

7.1.7.2 *EPA Method 17 Configuration*—For EPA Method 17 configuration, the sample filter is located in the duct and, therefore, naturally maintained at the flue gas temperature. The heating system is only required to maintain the probe and connecting umbilical cord to at least 120°C. If the flue gas temperature is less than 120°C, then EPA Method 5 configuration must be used.

7.1.8 *Condensing/Absorbing System*—consists of eight impingers immersed in an ice bath and connected in series with leak-free ground glass fittings or other noncontaminating leak-free fittings. (At no time is silicon grease or other greases to be used for this method). The first,

second, fourth, fifth, sixth, and eighth impingers are of the Greenburg–Smith design modified by replacing the standard tip with a 1.3-cm (0.5-in.)-ID straight glass tube extending to about 1.3 cm (0.5 in.) from the bottom of the flask. The third and seventh impingers are also Greenburg–Smith design, but with the standard tip including the glass impinging plate. The first, second, and third impingers contain aqueous 1 N potassium chloride (KCl) solution. The fourth impinger contains an aqueous solution of 5% $^{V}/_{V}$ nitric acid (HNO₃) and 10% $^{V}/_{V}$ hydrogen peroxide (H₂O₂). The fifth, sixth, and seventh impingers contain an aqueous solution of 4% $^{W}/_{V}$ potassium permanganate (KMnO₄) and 10% $^{V}/_{V}$ sulfuric acid (H₂SO₄). The last impinger contains silica gel or an equivalent desiccant. Refer to Note 1.

Note 1—When flue gas streams are sampled with high moisture content (>20%), additional steps must be taken to eliminate carryover of impinger contents from one sample type to the next. These steps must include use of oversized impinger(s) or use of an empty impinger between the KCl and $HNO_3-H_2O_2$. If a dry impinger is used, it must be rinsed as discussed in Section 13.2 of this method and the rinse added to the preceding impinger.

7.1.9 *Metering System*—vacuum gauge, leak-free pump, thermometers capable of measuring temperature to within $3^{\circ}C$ (5.4°F), and a dry gas meter or controlled orifice capable of measuring volume to within 2%.

7.1.10 *Barometer*— barometer capable of measuring atmospheric pressure to within 0.33 kPa (0.1 in. Hg). In many cases, the barometric reading may be obtained from a nearby National Weather Service station, in which case, the station value (which is the absolute barometric pressure) shall be requested. An adjustment for elevation differences between the weather station and sampling point shall be applied at a rate of negative 0.33 kPa (0.1 in. Hg) per 30 m (100 ft) elevation increase or vice versa for elevation decrease.

7.1.11 *Gas Density Determination Equipment*—temperature sensor and pressure gauge, as described in Section 2.3 and 2.4 of EPA Method 2. The temperature sensor shall, preferably, be permanently attached to the pitot tube or sampling probe in a fixed configuration, such that the sensor tip extends beyond the leading edge of the probe sheath and does not touch any metal. Alternative temperature sensor configurations are described in Section 2.1.10 of EPA Method 5. If necessary, a gas analyzer can be used to determine dry moleculer weight of the gas (refer to EPA Method 3).

7.2 Digestion Apparatus

7.2.1 *Dry Block Heater or Hot Water Bath*—a heater capable of maintaining a temperature of 95°C is required for digestion of samples, similar to that described in EPA SW846 Method 7470.

7.2.2 Ice Bath

7.2.3 *Digestion Flasks*—Use 50- to 70-mL tubes or flasks with screw caps that will fit a dry block heater. For a water bath, 300-mL biological oxygen demand bottles for SW846 Method

7470 are to be used. In addition, borosilicate glass test tubes, 35- to 50-mL volume, with rack are needed.

7.2.4 *Microwave or Convection Oven and PTFE Digestion Vessels*—120 mL, or equivalent digestion vessels with caps equipped with pressure relief valves for the dissolution of ash, along with a capping station or the equivalent to seal the digestion vessel caps. Use a vented microwave or convection oven for heating. In addition, polymethylpentene (PMP) or equivalent volumetric flasks are recommended for the digested ash solutions.

7.3 *Analytical Equipment*—dedicated mercury analyzer or equivalent apparatus for the analysis of mercury via CVAAS. Alternatively, CVAFS may be used. CVAAS is a method based on the absorption of radiation at 253.7 nm by mercury vapor. The mercury is reduced to the elemental state and aerated from solution in a closed system. The mercury vapor passes through a cell positioned in the light path of an atomic absorption spectrometer. Absorbency is measured as a function of mercury concentration. A soda-lime trap and a magnesium perchlorate trap must be used to precondition the gas before it enters the absorption cell.

8. Reagents and Materials

8.1 *Purity of Reagents*—Reagent-grade chemicals shall be used in all tests. Unless otherwise indicated, it is intended that all reagents conform to the specifications of the Committee on Analytical Reagents of the American Chemical Society, where such specifications are available.⁶ Other grades may be used, provided it is first ascertained that the reagent is of sufficiently high purity to permit its use without lessening the accuracy of the determination.

8.2 *Purity of Water*—Unless otherwise indicated, references to water shall be understood to mean reagent water as defined by Type II in ASTM Specification D 1193.

8.3 Reagents:

8.3.1 *Boric Acid* (H_3BO_3) —purified reagent grade.

8.3.2 *Hydrochloric Acid (HCl)*—trace metal-grade concentrated hydrochloric acid, with a specific gravity of 1.18.

8.3.3 *Hydrofluoric Acid (HF)*—concentrated hydrofluoric acid, 48%–50%.

8.3.4 *Hydrogen Peroxide* (H_2O_2) —30%^V/_v hydrogen peroxide.

8.3.5 *Hydroxylamine Sulfate* $(NH_2OH \cdot H_2SO_4)$ —solid.

⁶ "Reagent Chemicals, American Chemical Society Specifications," Am. Chemical Soc., Washington, DC. For suggestions on the testing of reagents not listed by the American Chemical Society, see "Reagent Chemicals and Standards," by Joseph Rosin, D. Van Nostrand Co., Inc., New York, NY, and the "United States Pharmacopeia."

8.3.6 *Hydroxylamine Hydrochloride* ($NH_2OH \cdot HCl$)—10% solution

8.3.6 *Sodium Chloride (NaCl)*—solid.

8.3.7 *Mercury Standard Solution*—a certified (1000 µg/mL) mercury standard.

8.3.7 *Nitric Acid* (HNO_3)—trace metal-grade concentrated nitric acid with a specific gravity of 1.42.

8.3.8 *Potassium Chloride (KCl)*—solid.

8.3.9 *Potassium Permanganate (KMnO₄)*—solid.

8.3.10 *Potassium Persulfate* $(K_2S_2O_8)$ —solid.

8.3.11 Stannous Chloride $(SnCl_2 \cdot 2H_2O)$ —solid.

8.3.12 Sulfuric Acid (H_2SO_4) —trace metal-grade concentrated sulfuric acid, with a specific gravity of 1.84.

8.4 *Materials:*

8.4.1 *Indicating Silica Gel*—with a size of 6–16 mesh.

8.4.2 Crushed or Cubed Ice.

8.4.3 Sample Filters—quartz fiber filters, without organic binders, exhibiting at least 99.95% efficiency (<0.05% penetration) for 0.3- μ m dioctyl phthalate smoke particles and containing less than 0.2 μ g/m² of mercury. Test data provided by filter manufacturers and suppliers stating filter efficiency and mercury content are acceptable. However, if no such results are available, determine filter efficiency using ASTM Test Method D 2986, and analyze filter blanks for mercury prior to emission testing. Filter material must be unreactive to sulfur dioxide (SO₂) or sulfur trioxide (SO₃).⁷

8.4.4 *Filter Papers*—for filtration of digested samples. The filter paper must have a particle retention of $>20 \mu m$ and filtration speed of >12 sec.

8.4.5 Nitrogen Gas (N_2) —carrier gas of at least 99.998% purity. Alternatively, argon gas may be used.

⁷ Felix, L.G.; Clinard, G.I.; Lacey, G.E.; McCain, J.D. "Inertial Cascade Impactor Substrate Media for Flue Gas Sampling," U.S. Environmental Protection Agency, Research Triangle Park, NC 27711, Publication No. EPA-600/7-77-060; June 1977, 83 p.

8.4.6 *Soda Lime*—indicating 4- to 8-mesh absorbent for trapping carbon dioxide.

8.4.7 Sample Containers—glass with PTFE-lined lids.

8.5 Sampling Reagents

8.5.1 *KCl Absorbing Solution* (1 mol/L)—Dissolve 74.56 g of KCl in 500 mL of reagent water in a 1000-mL volumetric flask, swirl to mix, and dilute to volume with water. Mix well. A new batch of solution must made prior to each field test.

8.5.2 *HNO*₃–*H*₂*O*₂ *Absorbing Solution* (5%^V/_V HNO₃, 10%^V/_V H₂O₂)—Add slowly, with stirring, 50 mL of concentrated HNO₃ to a 1000-mL volumetric flask containing approximately 500 mL of water, and then add carefully, with stirring, 333 mL of 30%^V/_V H₂O₂. Dilute to volume with water. Mix well. A new batch of solution must made prior to each field test.

8.5.3 H_2SO_4 -*KMnO₄ Absorbing Solution* (4%^W/_V KMnO₄, 10%^V/_V H₂SO₄)—Mix carefully, with stirring, 100 mL of concentrated H₂SO₄ into approximately 800 mL of water. When mixing, be sure to follow standard acid to water addition procedures and safety precautions associated with strong acids. Then add water, with stirring, to make 1 L. This solution is 10%^V/_V H₂SO₄. Dissolve, with stirring, 40 g of KMnO₄ into 10%^V/_V H₂SO₄, and add 10%^V/_V H₂SO₄, with stirring, to make 1 L. (Warning—See 9.1.1). H₂SO₄–KMnO₄ absorbing Solution must be made daily.

8.6 Rinse Solutions for Sample Train

8.6.1 0.1 N HNO_3 Solution—A certified reagent grade 0.1 N HNO₃ solution can be purchased directly or can be made by slowly adding 12.5 mL of concentrated HNO₃ to a 2000-mL volumetric flask containing approximately 500 mL of water, then diluting with water to volume.

8.6.2 $10\%\% \ W_V HNO_3$ Solution—Mix carefully, with stirring, 100 mL of concentrated HNO₃ into approximately 800 mL of water. When mixing, be sure to follow standard acid to water addition procedures and safety precautions associated with strong acids. Then add water, with stirring, to make 1 L.

8.6.3 $10\% W_{V}$ Hydroxylamine solution—Add 100 g Hydroxylamine sulfate and 100 grams sodium chloride to a 1000-mL volumetric flask containing approximately 500 mL of water. After the Hydroxylamine sulfate and sodium chloride has been dissolved, dilute with water to volume. As an alternative a 10% hydroxylamine hydrochloride solution can be used in all cases as a replacement for the hydroxylamine sulfate/sodium chloride solution.

8.7 Sample Digestion Reagents:

8.7.1 Boric Acid Solution $(4\%^{W}/_{V})$ —Dissolve 4 g H₃BO₃ in water, and dilute to 100 mL.

8.7.2 Aqua Regia ($HCl:HNO_3$ 3:1)—Add 3 parts concentrated HCl to 1 part concentrated HNO₃. Note that this should be made up in advance and allowed to form a dark orange color. This mixture should be loosely capped, as pressure will build as gases form.

8.7.3 Saturated Potassium Permanganate Solution $(5\%^W/_V)$ —Mix 5 g KMnO₄ into water, dilute to 100 mL, and stir vigorously.

8.7.4 *Potassium Persulfate Solution* $(5\%^W/_V)$ —Dissolve 5 g K₂S₂O₈ in water, and dilute to 100 mL.

8.8 Analytical Reagents:

8.8.1 *Hydrochloric Acid Solution* $(10\%^{V}/_{V})$ —Add 100 mL concentrated HCl to water, and dilute to 1 L. Be sure to follow all safety precautions for using strong acids.

8.8.2 Stannous Chloride Solution $(10\%^{V}/_{V})$ -Dissolve 100 g in $10\%^{V}/_{V}$ HCl, and dilute with $10\%^{V}/_{V}$ HCl to 1 L. Difficulty in dissolving the stannous chloride can be overcome by dissolving in a more concentrated HCl solution (such as 100 mL of $50\%^{V}/_{V}$ HCl) and diluting to 1 L with water. Note that care must be taken when adding water to a strong acid solution. Add a lump of mossy tin (~0.5 g) to this solution.

8.9 Mercury Standards:

8.9.1 10 mg/L Hg Stock Solution—Dilute 1 mL of 1000 mg/L Hg standard solution to 100 mL with $10\%^{\rm V}_{\rm V}$ HCl.

8.9.2 100 μ g/L Hg Stock Solution—Dilute 1 mL of 10 mg/L Hg stock solution to 100 mL with 10% $^{\rm V}/_{\rm V}$ HCl.

8.9.3 *Working Hg Standards*—Prepare working standards of 1.0, 5.0, 10.0, and 20.0 μ g/L Hg from the 100- μ g/L stock solution by diluting 1, 5, 10, and 20 mL each to 100 mL with 10% $^{V}/_{V}$ HCl.

Note 1—If samples to be analyzed are less than 1.0 μ g/L Hg, working standards should be prepared at 0.05, 0.1, 0.5, and 1.0 μ g/L Hg from a 10- μ g/L Hg standard solution.

8.9.4 *Quality Control Standard (QC)*—A quality control standard is prepared from a separate Hg standard solution. The QC standard should be prepared at a concentration of approximately one-half the calibration range.

8.10 *Glassware Cleaning Reagents*—Prior to any fieldwork, all glassware should be cleaned according to the guidelines outlined in EPA Method 29, section 5.1.1

9. Hazards

9.1 Warning:

9.1.1 Pressure may build up in the solution storage bottle because of a potential reaction between potassium permanganate and acid. Therefore, these bottles should not be fully filled and should be vented to relieve excess pressure and prevent explosion. Venting must be in a manner that will not allow contamination of the solution.

9.1.2 Hazards to personnel exist in the operation of the cold-vapor atomic absorption spectrophotometer. Refer to the manufacturer's instruction manual before operating the instrument.

9.1.3 Sample digestion with hot concentrated acids creates a safety problem. Observe appropriate laboratory procedures for working with concentrated acids.

9.2 Precaution:

9.2.1 The determination of microquantities of mercury species requires meticulous attention to detail. Good precision is generally unattainable without a high level of experience with stack-sampling procedures. Precision may be improved by knowledge of, and close adherence to, the suggestions that follow.

9.2.1.1 All glassware used in the method must be cleaned thoroughly prior to use in the field, as described in Section 8.10 of this method.

9.2.1.2 Use the same reagents and solutions in the same quantities for a group of determinations and the corresponding solution blank. When a new reagent is prepared or a new stock of filters is used, a new blank must taken and analyzed.

10. Sampling

10.1 Preparation for Test:

10.1.1 *Preliminary Stack Measurements*—Select the sampling site, and determine the number of sampling points, stack pressure, temperature, moisture, dry molecular weight, and range of velocity head in accordance with procedures of ASTM Test Method D 3154 or EPA Methods 1 through 4.

10.1.2 Select the correct nozzle diameter to maintain isokinetic sampling rates based on the range of velocity heads determined in 10.1.1.

10.1.3 Ensure that the proper differential pressure gauge is selected for the range of velocity heads (refer to EPA Method 2, Section 2.2).

10.1.4 It is suggested that an EPA Method 17 configuration be used; however, if an EPA Method 5 setup is to be used, then select a suitable probe length such that all traverse points can be sampled. Consider sampling from opposite sides of the stack to minimize probe length when a large duct or stack is sampled.

10.1.5 *Sampling Time and Volume*—The total sampling time for this method should be at least 2 but not more than 3 hours. Use a nozzle size that will guarantee an isokinetic gas sample volume between 1.0 dry cubic meters corrected to standard conditions (dscm) and 2.5 dscm. If traverse sampling is done (necessary for sampling at electric utilities), use the same points for sampling that were used for the velocity traverse as stated in Section 10.1.1 of this method. Each traverse point must be sampled for a minimum of 5 minutes.

11. Preparation of Apparatus

11.1 Pretest Preparation:

11.1.1 Weigh several 200- to 300-g portions of silica gel in airtight containers to the nearest 0.5 g. Record the total mass of the silica gel plus container on each container. Alternatively, the silica gel can be weighed directly in the impinger immediately prior to the train being assembled.

11.1.2 Desiccate the sample filters at $20^{\circ} \pm 5.6^{\circ}$ C ($68^{\circ} \pm 10^{\circ}$ F) and ambient pressure for 24 to 36 hours, weigh at intervals of at least 6 hours to a constant mass (i.e., <0.5-mg change from previous weighing), and record results to the nearest 0.1 mg. Alternatively, the filters may be oven-dried at 105°C (220°F) for 2 to 3 hours, desiccated for 2 hours, and weighed.

11.1.3 Clean all sampling train glassware as described in Section 8.10 before each series of tests at a single source. Until the sampling train is assembled for sampling, cover all glassware openings where contamination can occur.

11.2 Preparation of Sampling Train:

11.2.1 Assemble the sampling train as shown in Figure 1.

11.2.2 Place 100 mL of the KCl solution (see Section 8.5.1 of this method) in each of the first, second, and third impingers, as indicated in Figure 1.

11.2.3 Place 100 mL of the $HNO_3-H_2O_2$ solution (Section 8.5.2 of this method) in the fourth impinger, as indicated in Figure 1.

11.2.4 Place 100 mL of the H_2SO_4 -KMnO₄ absorbing solution (see Section 8.5.3 of this method) in each of the fifth, sixth, and seventh impingers, as indicated in Figure 1.

11.2.5 Transfer approximately 200 to 300 g of silica gel from its container to the last impinger, as indicated in Figure 1.

11.2.6 Prior to final train assembly, weigh and record the mass of each impinger. This information is required to calculate the moisture content of the sampled flue gas.

11.2.7 To ensure leak-free sampling train connections and to prevent possible sample contamination problems, use PTFE tape, PTFE-coated O-rings, or other noncontaminating material.

11.2.8 Place a weighed filter in the filter holder using a tweezer or clean disposable surgical gloves.

11.2.9 Install the selected nozzle using a noncontaminating rubber-type O-ring or equivalent when stack temperatures are less than 260° C (500° F) and an alternative gasket material when temperatures are higher. Other connecting systems, such as PTFE ferrules or ground glass joints, may also be used on the probe and nozzle.

11.2.10 Mark the probe with heat-resistant tape or by some other method to denote the proper distance into the stack or duct for each sampling point.

11.2.11 Place crushed or cubed ice around the impingers.

11.2.12 *Leak-Check Procedures*. Follow the leak-check procedures given in Section 4.1.4.1 (Pretest Leak Check), Section 4.1.4.2 (Leak Checks During the Sample Run), and Section 4.1.4.3 (Posttest Leak Checks) of EPA Method 5 or 17.

Note 2—If the flue gas temperature at the sampling location is greater than 260°C (above the temperature where PTFE or rubber-type seals can be used), the posttest leak check is determined beginning at the front end of the probe (does not include nozzle or sample filter holder for EPA Method 17).

12. Calibration and Standardization

12.1 Sampling Train Calibration:

12.1.1 *Probe Nozzle*—Refer to Sections 2.1.1 of either EPA Method 5 or 17.

- 12.1.2 *Pitot Tube*—Refer to Section 4 of EPA Method 2.
- 12.1.3 *Metering System*—Refer to Section 5.3 of either EPA Method 5 or 17.
- 12.1.4 *Probe Heater*—Refer to Section 7.1.7.1 and 7.1.7.2 of this method.
- 12.1.5 *Temperature Gauges* Refer to Section 4.3 of EPA Method 2.

12.1.6 *Leak Check of the Metering System*—Refer to Section 5.6 of EPA Method 5 or Section 5.5 of EPA Method 17.

12.1.7 *Barometer*—Calibrate the barometer to be used against a mercury barometer.

12.2 Atomic Absorption or Atomic Fluorescence Spectrometer Calibration—Perform instrument setup and optimization according to the manufacturer's specifications. Cold-vapor generation of mercury is performed via addition of stannous chloride solution to reduce oxidized mercury to its elemental state. The mercury-laden solution is then purged with a carrier gas into the atomic absorption cell. This procedure is used to calibrate the instrument using $10\%^{V}/_{V}$ HCl as the blank along with the standards described in Section 8.9.3. Calibration is verified by analyzing the QC standard prepared according to Section 8.9.4 of this method.

13. Procedures

13.1 Sampling Train Operation:

13.1.1 Maintain an isokinetic sampling rate within 10% of true isokinetic. For an EPA Method 5 configuration, maintain sample filter exit gas stream temperatures and probe within $\pm 15^{\circ}$ C of the flue gas temperature at the sampling location. However, at no time, regardless of the sample configuration, must the sample filter, probe, or connecting umbilical cord temperature be lower than 120° C.

13.1.2 Record the data, as indicated in Figure 2, at least once at each sample point but not less than once every 5 minutes.

13.1.3 Record the dry gas meter reading at the beginning of a sampling run, the beginning and end of each sampling time increment, before and after each leak check, and when sampling is halted.

13.1.4 Level and zero the manometer. Periodically check the manometer level and zero, because it may drift during the test period.

13.1.5 Clean the port holes prior to the sampling run.

13.1.6 Remove the nozzle cap. Verify that the filter and probe heating systems are up to temperature and that the pitot tube and probe are properly positioned.

Note 3—For an EPA Method 5 configuration, prior to starting the gas flow through the system, the sample filter exit gas temperature may not be at the hot box temperature. However, if the system is set up correctly, once flow is established, the sample filter exit gas temperature will quickly come to equilibrium.

13.1.7 Start the pump. Position the nozzle at the first traverse point with the nozzle tip pointing in the direction of flow. Seal the openings around the probe and port hole to prevent unrepresentative dilution of the gas stream. Read the pitot tube manometer, start the stopwatch, open and adjust the control value until the isokinetic sampling rate is obtained (refer to Section

4.1.5 from either EPA Method 5 or 17 for information on isokinetic sampling rate computations), and maintain the isokinetic rate at all points throughout the sampling period.

13.1.8 When sampling at one traverse point has been completed, move the probe to the next traverse point as quickly as possible. Close the coarse adjust valve, and shut the pump off when transferring the probe from one sample port to another. Exclude the time required to transfer the probe from one port to another from the total sampling time.

13.1.9 Traverse the stack cross section, as required by EPA Method 1.

13.1.10 During sampling, periodically check and, if necessary, adjust the probe and filter exit sample gas temperatures, as well as the zero of the manometer.

13.1.11 Add more ice, if necessary, to maintain a temperature of $<20^{\circ}C$ (68°F) at the condenser/silica gel outlet.

13.1.12 Replace the filter assembly if the pressure drop across the filter becomes such that maintaining isokinetic sampling is no longer possible. Conduct a leak check (refer to EPA Method 5 or 17, Section 4.1.4.2) before installing a new filter assembly. The total particulate mass and determination of particle-bound mercury includes all filter assembly catches.

13.1.13 In the unlikely event depletion of $KMnO_4$ via reduction reactions with flue gas constituents other than elemental mercury occurs, it may render it impossible to sample for the desired minimum time. This problem is indicated by the complete bleaching of the purple color of the acidified permanganate solution. If the purple color is lost in the first two H_2SO_4 –KMnO_4 impingers, then the sample must be repeated. If the gas stream is known to contain large amounts of reducing constituents (i.e., >2500 ppm SO₂) or breakthrough has occurred in previous sampling runs, then the following modification is suggested: the amount of $HNO_3-H_2O_2$ ($10\%^V/_V$) in the fourth impinger should be doubled, and/or a second $HNO_3-H_2O_2$ impinger should be used to increase the oxidation capacity for reducing gas components prior to the H_2SO_4 –KMnO₄ impingers.

13.1.14 Use a single train for the entire sample run, except when simultaneous sampling is required in two or more separate ducts or at two or more different locations within the same duct or when equipment failure necessitates a change of trains.

13.1.15 At the end of a sample run, turn off the coarse adjust valve, remove the probe and nozzle from the stack, record the final dry gas meter reading, and conduct a posttest leak check, as described in Section 4.1.4.3 of EPA Method 5. Also, leak-check the Pitot lines as described in EPA Method 2, Section 3.1. The lines must pass the leak check to validate the velocity head data.

13.1.16 Calculate percent isokinetic to determine whether the run was valid or another test run should be performed (refer to EPA Method 5 or 17).

13.2 Sample Recovery:

13.2.1 Allow the probe to cool before proceeding with sample recovery. When the probe can be safely handled, wipe off all external particulate matter near the tip of the probe nozzle, and place a rinsed, noncontaminating cap over the probe nozzle to prevent losing or gaining particulate matter. Do not cap the probe tip tightly while the sampling train is cooling; a vacuum can form in the filter holder, with the undesired result of drawing liquid from the impingers onto the filter.

13.2.2 Before moving the sampling train to the cleanup site, remove the probe from the sampling train, and cap the open outlet. Be careful not to lose any condensate that may be present. Cap the filter inlet where the probe was fastened. Remove the umbilical cord from the last impinger, and cap the impinger. Cap the filter holder outlet and impinger inlet. Use noncontaminating caps, such as ground-glass stoppers, plastic caps, serum caps, or PTFE tape, to close these openings.

13.2.3 Alternatively, the following procedure may be used to disassemble the train before the probe and filter holder/oven are completely cooled. Initially disconnect the filter holder outlet/impinger inlet, and loosely cap the open ends. Then disconnect the probe from the filter holder or cyclone inlet, and loosely cap the open ends. Cap the probe tip, and remove the umbilical cord as previously described.

13.2.4 Transfer the probe and filter–impinger assembly to a clean area that is protected from the wind and other potential causes of contamination or loss of sample. Inspect the train before and during disassembly, and note any abnormal conditions.

13.2.5 The impinger train sample recovery scheme is illustrated in Figure 3.

13.2.6 *Container 1 (Sample Filter)*—Carefully remove the sample filter from the filter holder so as not to lose any ash, weigh filter and ash, and place the filter in a labeled petri dish container. To handle the filter, use either acid-washed polypropylene or PTFE-coated tweezers or clean, disposable surgical gloves rinsed with water and dried. If it is necessary to fold the filter, make certain the particulate cake is inside the fold. Transfer any particulate matter or filter fibers that adhere to the filter holder gasket to the filter in the petri dish. A dry (acid-cleaned) nonmetallic bristle brush should be used to remove any remaining particulate matter. Do not use any metal-containing materials when recovering this train. Immediately cover and seal the labeled petri dish.

13.2.7 Container 2/2a (All Rinses in Front of the Sample Filter)

13.2.7.1 Case 1: Includes Gravimetric Particulate Determination in Addition to Mercury

Quantitatively recover particulate matter and any condensate from all components prior to the sample filter. A nonmetallic brush may be used for removing particulate matter. All front-half components (all components prior to the sample filter) are then rinsed with acetone as outlined in

EPA Method 5 or 17. The acetone rinse is then placed into a container (Container 2a) for which the tare weight has been recorded. Container 2a, with a ribbed watch glass over the top, is placed in a fume hood until the acetone has completely evaporated. After the front-half components have been rinsed with acetone, then rinse these components with 0.1 N HNO₃. The 0.1 N HNO₃ rinse is placed in Container 2.

13.2.7.2 Case 2: Mercury Determination Only (No Acetone Rinse)

Quantitatively recover particulate matter and any condensate from all components prior to the sample filter. A nonmetallic brush may be used for removing particulate matter. The front-half components are then rinsed with 0.1 N HNO₃, and this rinse is placed in Container 2.

13.2.8 Container 3 (Impingers 1 through 3, KCl Impinger Contents and Rinses):

13.2.8.1 Dry the exterior surfaces of Impingers 1, 2, and 3. Then weigh and record the mass of each impinger (to the nearest 0.5 g).

13.2.8.2 Clean the filter support, the back half of the filter housing, and connecting glassware by thoroughly rinsing with 0.1 N HNO_3 . Pour the rinse into a glass sample Container 3.

13.2.8.3 Carefully add small amounts of 5% $^{W}/_{V}$ KMnO₄ solution very slowly to each KCl impinger and gently mix the impinger solution. Continue adding KMnO₄ solution until a purple color is obtained. Let the impingers sit for approximately 15 minutes to ensure the purple color persists.

13.2.8.4 Pour all of the liquid from the three KCl impingers into Container 3.

13.2.8.5 Rinse the impingers and connecting glassware with $10\%^{V}/_{V}$ HNO₃. Although unlikely, if deposits remain on the impinger surfaces, remove them by doing another $10\%^{V}/_{V}$ HNO₃ rinse that has a very small amount (several drops) of $10\%^{W}/_{V}$ hydroxylamine solution added to the HNO₃ rinse solution. Rinse each of the KCl impingers with this solution until the brown stains are removed. Add these rinses to Container 3. If the solution in Container 3 becomes clear, add a small amount of the 5% $^{W}/_{V}$ KMnO₄ solution until a pink or slightly purple color is obtained. Check again after 90 min to ensure the purple color remains.

13.2.8.6 Perform a final rinse of the impingers and connecting glassware with 0.1 N HNO_3 , and add to Container 3.

13.2.8.7 Do a final rinse of all glass components with water which is discarded.

13.2.8.8 Mark the height of the fluid level in Container 3, seal, and clearly label the contents.

13.2.9 Container 4 (Impinger 4, HNO₃-H₂O₂ Impinger Contents and Rinses):

13.2.9.1 Dry the exterior surfaces of Impinger 4. Then weigh and record the mass of this impinger (to the nearest 0.5 g).

13.2.9.1 Pour the $HNO_3-H_2O_2$ absorbing solution into sample Container 4.

13.2.9.2 Rinse the H_2O_2 -HNO₃ impinger and connecting glassware a minimum of two times with 0.1 N HNO₃, and pour the rinses into Container 4. Do a final rinse with water and discard water.

13.2.10 Container 5 (Impingers 5 through 7, H_2SO_4 -KMn O_4 Impinger Contents and Rinses):

13.2.10.1 Dry the exterior surfaces of Impingers 5, 6, and 7. Then weigh and record the mass of each impinger (to the nearest 0.5 g).

13.2.10.2 Pour all of the liquid from the three H_2SO_4 -KMnO₄ impingers into a glass sample Container 5.

13.2.10.3 Rinse the H_2SO_4 -KMnO₄ impingers and connecting glassware a minimum of two times with 0.1 N HNO₃, and pour the rinses into Container 5. A third rinse must then be done (this rinse will remove any brown stains from the surface of the impingers). This rinse consists of 0.1N HNO₃ and several drops of 10% ^w/_v hydroxylamine solution (either the NH₂OH/NaCl solution or the NH₂OH·HCl solution). This rinse must have enough 10% ^w/_v hydroxylamine solution such that the brown stains are easily removed. If they are not easily removed add several more drops of 10% ^w/_v hydroxylamine solution until the stains are completely gone. Add this rinse to Container 5. If the solution in Container 5 becomes clear, add small amounts of the H₂SO₄-KMnO₄ solution until a pink or slightly purple color is obtained.

13.2.10.4 Perform a final 0.1 N HNO_3 rinse of the impingers and connecting glassware follow by a water rinse. The 0.1 N HNO_3 rinse is added to Container 5, and the water rinse is discarded.

13.2.10.5 Mark the height of the fluid level, seal the container, and clearly label the contents.

Note 4—As stated earlier in the warning in Section 9.1.1, pressure can build up in the sample storage flask because of the potential reaction of $KMnO_4$ with acid. Do not fill the container completely, and take precautions to relieve excess pressure.

13.2.11 Container 6 (Impinger 8, Silica Gel Impinger Contents):

13.2.11.1 Dry the exterior surfaces of Impinger 8. Then weigh and record the mass of this impinger (to the nearest 0.5 g).

13.2.11.2 Note the color of the indicating silica gel to determine whether it has been completely spent, and make a notation of its condition. If spent, the silica gel must be either regenerated or disposed of.

13.2.12 Solution Blanks (Containers 7–11)

Solution blanks are taken each time new reagents are prepared. *Note:* The amount of solution collected for the blanks stated below is a suggested volume.

13.2.12.1 *Container 7 (0.1 N HNO*₃ *Blank)*—Place 50 mL of the 0.1 N HNO₃ solution used in the sample recovery process into a properly labeled container. Seal the container.

13.2.12.2 *Container 8 (1 N KCl Blank)*—Place 50 mL of the 1 N KCl solution used as the impinger solution into a properly labeled container. Seal the container.

13.2.12.3 Container 9 $(5\%^{V}/_{V} HNO_{3}-10\%^{V}/_{V} H_{2}O_{2} Blank)$ —Place 50 mL of the HNO₃-H₂O₂ solution used as the nitric acid impinger reagent into a properly labeled container. Seal the container.

13.2.12.4 Container 10 (H_2SO_4 -KMnO_4 Blank)—Place 50 mL of the H_2SO_4 -KMnO_4 solution used as the impinger solution in the sample recovery process into a properly labeled container. Refer to **Note 4** in Section 13.2.10.5 of this method.

13.2.12.5 *Container 11 (10%* $^{W}/_{V}$ *Hydroxylamine Solution)*—Place 100 mL of hydroxylamine solution into a properly labeled sample container. Seal the container.

13.2.13 *Container 12 (Sample Filter Blank)*—Once during each field test, place into a properly labeled petri dish three unused blank filters from the same lot as the sampling filters. Seal the petri dish.

13.2.14 After all of the samples have been recovered, they must be analyzed within 45 days.

13.2.15 After all impingers and connectors have been properly rinsed and the solutions recovered, the glassware should be cleaned according to the procedures in Section 8.10 or triple-rinsed with $10\%^{V}_{V}$ HNO₃ followed by a rinsing with water. If a new source is to be sampled or if there are any brown stains on the glassware, then the glassware must be cleaned according to procedures in Section 8.10 of this method. If multiple sites are to sampled during a single mobilization, an exception to this procedure will be allowed. In this case, a triple rinsing of the glassware with $10\%^{V}_{V}$ HNO₃ solution followed by a water rinse prior to sampling can be used as an alternative to the procedures in Section 8.10. However, if there are any brown stains on the glassware, then the glassware must be cleaned according to procedures in Section 8.10 of this method.

13.3 Sample Preparation:

13.3.1 Ash Sample (Containers 1 and 2)

13.3.1.1 *Case 1: Includes Gravimetric Particulate Determination in Addition to Mercury*—The gravimetric particulate loading is determined from the mass of the ash on the filter (Container 1) and the residual particulate from the acetone rinse (Container 2a), as outlined in EPA Method 5 or 17. If a large amount of ash is on the filter, carefully remove the ash to create a raw ash sample from which a representative weighed aliquot can be taken for digestion. If the mass of ash collected on the filter is small (less than 0.5 g), digest the entire filter along with the ash. Dissolve the residual particulate from Container 2a using concentrated HNO₃. This solution is then added to Container 2 (0.1 N HNO₃ probe rinse). The ash material from Container 1 is then digested using the procedures described in Section 13.3.2 of this method. The same procedure is used to determine the mercury on the sample filter blank.

Use a modification of EPA SW 846 7470 to digest the sample in Container 2 prior to analysis. The main modification is that the volumes of reagents and sample have been reduced tenfold to reduce waste. This reduction in reagent volume is acceptable because modern dedicated mercury analyzers do not require the large volumes that previous manual methods required. Transfer a 10-mL aliquot of the sample to a digestion tube with a screw cap.

13.3.1.2 *Case 2: Mercury Determination Only*—The same procedures are followed as described previously in Section 13.3.1.1 with the exception that there is no Container 2a.

13.3.2 *Ash Digestion*—Accomplish the complete dissolution of ash by one of the following methods or an equivalent alternative method. The following methods are for the dissolution of inorganic samples, such as ash or sediments, when an analysis of trace elements including mercury is done.

13.3.2.1 *Microwave Digestion*—The use of this method assumes proper training in microwave digestion techniques. In addition, this method is tailored for a CEM (continuous emission monitor) microwave digestion system. A 0.5-g ash sample, accurately weighed to 0.0001 g, is placed in a PTFE microwave digestion vessel with 3 mL of concentrated HF, 3 mL of concentrated HNO₃, and 3 mL of concentrated HCl. The vessel is sealed and placed in the microwave (along with other vessels). The vessels are slowly heated to a pressure of 347 kPa (50 psi), which is held for 5 minutes, followed by heating to a pressure of 550 kPa (80 psi), which is held for 20 minutes. The vessels are allowed to cool to room temperature before venting. 15 mL of $4\%^{W/_V}$ boric acid is added to each vessel. The vessels are sealed and placed in the microwave again. The vessels are slowly heated back to a pressure of 347 kPa (50 psi) and held for 10 minutes. The vessels are again allowed to cool to room temperature before venting. The contents of each vessel are quantitatively transferred to a 50-mL PMP or polypropylene (PP) volumetric flask and diluted; note that care must be taken in adding water to a strong acid solution.

13.3.2.2 *Conventional Digestion*—The use of this method assumes proper training in PTFE bomb digestion techniques. Place a 0.5-g ash sample, accurately weighed to 0.0001 g, in a PTFE

digestion vessel with 7 mL of concentrated HF and 5 mL of aqua regia. Seal the vessel, and place it in an oven or water bath at 90°C for a minimum of 8 hours (these may be heated overnight). Cool the vessel to room temperature before venting. Add 3.5 g of boric acid and 40 mL of water to each vessel. Seal the vessels, and place them in the oven or water bath for an additional 1 hour. Cool the vessels again to room temperature before venting. Quantitatively transfer the contents of each vessel to a 100-mL PMP, PP, or glass volumetric flask and dilute. Note that care must be taken in adding water to a strong acid solution.

Preparation of Aqueous KCl Impinger Solution (Containers 3 and 8)—Dilute 13.3.3 sample in a 500-mL volumetric flask to volume with water, and mix. Use a modification of EPA SW 846 7470 to digest the sample prior to analysis. The main modification is that the volumes of reagents and sample have been reduced tenfold to reduce waste. This reduction in reagent volume is acceptable because modern dedicated mercury analyzers do not require the large volumes that previous manual methods required. Transfer a 10-mL aliquot of the sample to a digestion tube with a screw cap. Add 0.5 mL of concentrated H₂SO₄, 0.25 mL of concentrated HNO₃, and 1.5 mL of 5% $^{W}/_{V}$ KMnO₄ solution. Mix the solution, and allow it to stand for 15 minutes. Add 0.75 mL of 5% $^{W}/_{V}$ K₂S₂O₈ solution, and loosely cap the tube. Place the tube in a dry block heater or water bath equipped with a temperature probe, and heat to 95°C. Do not allow the temperature to exceed 95°C. Hold the sample at 95°C for 2 hours before allowing it to cool to room temperature. The purple color from the added KMnO₄ solution must remain throughout the digestion. Clearing of the solution during the heating indicates the depletion of KMnO₄. If the solution goes clear add more KMnO₄ to the sample until a purple color persists. Prior to analysis, add 1 mL of 10%^W/_v hydroxylamine sulfate solution to the sample. The sample solution should remain clear after addition of hydroxylamine sulfate. Record the volumes of the solution additions used in the preparation procedure and adjust the DF factor in equation 9 as necessary.

13.3.4 Preparation of HNO₃-H₂O₂ Impinger Solution (Containers 4 and 9)—Dilute sample in a 250-mL volumetric flask to volume with water, and mix. Treat the sample with a modified version of EPA SW 846 7470. Modifications to the method are necessary to properly treat the H₂O₂-containing impinger solution before the analysis with CVAAS. The modifications include the addition of HCl, the use of an ice bath during the KMnO₄ addition, and the slow addition of the KMnO₄. Transfer a 5-mL aliquot of the sample to a digestion tube with a screw cap. Add 0.25 mL of concentrated HCl, 0.25 mL of concentrated H₂SO₄, place the tube in an ice bath, and allow it to cool for 15 minutes. The destruction of H₂O₂ is accomplished by slow addition of saturated KMnO₄ solution in 0.25-mL increments along the inside of the digestion tube. The violence of this reaction requires careful, slow addition of the KMnO₄ for safety reasons and to avoid loss of analyte. Cool the sample for 15 minutes in between each addition, and mix the sample prior to each addition. After the first five additions, increase the increments to 0.5 mL. Carry out the addition of KMnO₄ until the solution remains purple, indicating complete reaction of the H₂O₂. Record the volume of saturated KMnO₄ solution added to the sample. Add 0.75 mL of 5% $^{W}/_{v}$ K₂S₂O₈ solution to the sample, and then cap the tube loosely. Place the tubes in a dry block heater or water bath equipped with a temperature probe, and heat to 95°C. Do not allow the temperature to exceed 95°C. Maintain the sample at 95°C for 2 hours before allowing it to cool to room temperature. Note that the purple color due to KMnO₄ must remain throughout the digestion. Clearing of the solution during the heating indicates the depletion of KMnO₄. Before

doing the analysis, add 1mL $10\%^{W}/_{V}$ of hydroxylamine sulfate solution to the sample. The sample should then become clear. Record the volumes of the solution additions used in the preparation procedure and adjust the DF factor in equation 13 as necessary.

13.3.5 Preparation of H₂SO₄-KMnO₄ Impinger Solution (Containers 5 and 10)—Prepare the entire solution immediately prior to analysis. Dissolve by incrementally adding approximately 500 mg of solid hydroxylamine sulfate into the sample until a clear, colorless solution persists. (This is to ensure that a representative aliquot sample can be taken and that any mercury contained in the manganese dioxide that forms from the permanganate solution will be removed). Add the hydroxylamine slowly because of the violence of this reaction. Dilute the sample in a 500mL volumetric flask to volume with water, and mix. Transfer a 10-mL aliquot of the sample to a digestion tube with a screw cap. Add 0.75 mL of 5% $^{W}/_{V}$ K₂S₂O₈ solution and 0.5mL of concentrated HNO₃, and loosely cap the tube. Place the tube in a dry block heater or water bath equipped with a temperature probe, and heat to 95°C. Do not allow the temperature to exceed 95°C. Hold the sample at 95°C for 2 hours before allowing it to cool to room temperature. The purple color of the KMnO₄ solution must remain throughout the digestion. Clearing of the solution during the heating indicates the depletion of KMnO₄. Prior to analysis, add 1 mL of $10\%^{W/_{V}}$ of hydroxylamine sulfate solution to the sample. The sample solution should remain clear after addition of hydroxylamine sulfate. Record the volumes of the solution additions used in the preparation procedure and adjust the DF factor in equation 12 as necessary.

13.3.6 Simplification of the Digestion—If an acetone rinse was not used for gravimetric particulate determination or it is very clear, there is insignificant organic material present in the sampled gas stream; then the digestion procedure for the $HNO_3-H_2O_2$ and $H_2SO_4-KMnO_4$ impinger solutions may be simplified by omitting the persulfate digest. The persulfate digest is performed for the purpose of oxidizing certain organics. Because this method is specific to coal combustion systems where organic compounds are usually insignificant,⁸ this digest may be omitted because the H_2O_2 is sufficient to oxidize most compounds. The decision to omit this procedure should be made based on the gas stream being sampled and/or verification that organics resistant to H_2O_2 oxidation are not present. If unsure whether organics are present or if an acetone rinse has been used, then the total digestion procedure is required.

13.3.6.1 Simplified Procedure for the Preparation of $HNO_3-H_2O_2$ Impinger Solution—If the simplified procedure can be used for the $HNO_3-H_2O_2$ impinger solution, the concentrated H_2SO_4 and $5\%^W/_V K_2S_2O_8$ are not added to the $HNO_3-H_2O_2$ aliquot sample. Also it is not necessary to heat the sample to $95^{\circ}C$ followed by 2 hours of cooling. However, it is still necessary that the concentrated HCl be added to the solution.

⁸ "A Comprehensive Assessment of Toxic Emissions from Coal-Fired Power Plants: Phase I Results from the U.S. Department of Energy Study," Prepared for the U.S. Department of Energy Federal Energy Technology Center, Contract No. DE-FC21-93MC30097, Energy & Environmental Research Center, University of North Dakota, Grand Forks, ND, 1996.

Just before doing the analysis, add 1mL $10\%^{W}_{V}$ of hydroxylamine solution to the sample. The sample should then become clear. If the simplified procedure is used, $V(K_2S_2O_8)$ and $V(H_2SO_4)$ are zero when calculating DF in Equation 12 Section 15.

13.3.6.2 Simplified procedure for the Preparation of H_2SO_4 -KMnO₄ Impinger Solution—If the simplified procedure can be used for the H_2SO_4 -KMnO₄ impinger solution, the concentrated HNO₃ and 5% ^W/_V K₂S₂O₈ are not added to the H₂SO₄-KMnO₄ aliquot sample. Also it is not necessary to heat the sample to 95 °C followed by 2 hours of cooling. Just before doing the analysis, add 1mL 10% ^W/_V of hydroxylamine solution to the sample. The sample should then become clear. If the simplified procedure is used, V(K₂S₂O₈) and V(HNO₃) are zero when calculating DF in Equation 13 Section 15.

13.3.7 *Reagent blanks (Containers 8 through 10)*—These samples are <u>not</u> diluted prior to taking an aliquot. Once an aliquot is taken the preparation steps for each of the solutions (as well as the mercury concentration calculations) are the same as described above. These are Section 13.3.3 for the aqueous KCl reagent blank, Section 13.3.4 for the $HNO_3-H_2O_2$ reagent blank, and Section 13.3.6.2 for the H_2SO_4 -KMnO₄ reagent blank.

13.3.8 0.1 N HNO₃ and $10\%^{W}/_{V}$ Hydroxylamine Rinse Solutions (Containers 7 and 11)—These solutions can be analyzed directly for mercury without any preparation steps.

13.4 *Sample Analysis*—Analyze all of the prepared solutions by CVAAS or CVAFS following the guidelines specified by the instrument manufacturer.

13.4.1 *QA/QC*—For this method, it is important that both the sampling team and analytical people be very well trained in the procedures. This is a complicated method that requires a high-level of sampling and analytical experience. For the sampling portion of the QA/QC procedure, both solution and field blanks are required. It should be noted that if high-quality reagents are used and care is taken in their preparation and in the train assembly, there should be little, if any, mercury measured in either the solution or field blanks.

As stated in Section 13.2.12 of this method, solution blanks will be taken and analyzed every time a new batch of solution is prepared. If mercury is detected in these solution blanks, the concentration is subtracted from the measured sample results. The maximum amount that can be subtracted is 10% of the measured result or 10 times the detection limit of the instrument which ever is lower. If the solution blanks are greater than 10% the data must be flagged as suspect.

A field blank is performed by assembling an impinger train, transporting it to the sampling location during the sampling period, and recovering it as a regular sample. These data are used to ensure that there is no contamination as a result of the sampling activities. A minimum of one field blank at each sampling location must be completed for each test site. Any mercury detected in the field blanks cannot be subtracted from the results. Whether or not the mercury detected in the field blanks is significant is determined based on the QA/QC procedures established prior to the testing. At a minimum, if field blanks exceed 30% of the measured value at the corresponding location, the data must be flagged as suspect.

The QA/QC for the analytical portion of this method is that every sample, after it has been prepared, is to be analyzed in duplicate with every tenth sample analyzed in triplicate. These results must be within 10% of each other. If this is not the case, then the instrument must be recalibrated and the samples reanalyzed. In addition, after every ten samples, a known spike sample must be analyzed. For the ash samples, a certified reference ash sample (may be purchased from NIST) is to be digested and analyzed at least once during the test program. It is also suggested that the QA/QC procedures developed for a test program include submitting, on occasion, spiked mercury samples to the analytical laboratory by either the prime contractor if different from the laboratory or an independent organization.

14. Flue Gas Calculations

14.1 Dry Gas Volume—Calculate the dry gas sample volume, $V_{m(std)}$, at standard conditions using Equation 1.

$$V_{m(std)} = V_m Y\left(\frac{T_{std}}{T_m}\right) \left[\frac{P_{bar} + \Delta H}{P_{std}}\right] = K_1 V_m Y \frac{P_{bar} + \Delta H}{T_m}$$
[Eq. 1]

where:

 P_{bar} = barometric pressure at the sampling site, kPa (in. Hg) = standard absolute pressure, 101.3 kPa (29.92 in. Hg) P_{std} T_m = absolute average dry gas meter temperature (refer to Figure 2), K ($^{\circ}$ R) = standard absolute temperature, 293 K ($528^{\circ}R$) T_{std} = volume of gas sample as measured by dry gas meter, dcm (dscf) V_m = volume of gas sample measured by the dry gas meter, corrected to standard V_{m(std)} conditions, dscm (dscf) Y = dry gas meter calibration factor = average pressure differential across the orifice meter (refer to Figure 2), kPa (in. Hg) ΔH $= 2.894 \text{ K/kPa} (17.64^{\circ}\text{R/in. Hg})$ \mathbf{K}_1

Note 5—Equation 1 can be used as written unless the leakage rate observed during any of the mandatory leak checks (i.e., leak checks conducted prior to component changes or following the test) exceeds the maximum acceptable leakage rate, L_a , equal to 0.00057 m³/min (0.02 cfm) or 4% of the average sampling rate, whichever is less. If the leakage rate observed during the posttest leak check, L_p , or an individual leakage rate observed during the leak check conducted prior to the "ith" component change (I = 1, 2, 3, ...n), L_i , exceeds L_a , then Equation 1 must be modified as follows:

(a) **Case I**. No component changes made during sampling run. In this case, replace V_m with the expression:

$$[V_m - (L_p - L_a)\theta]$$

where:

 L_p = leakage rate observed during the posttest leak check, m³/min (cfm)

- $L_a = maximum$ acceptable leakage rate for either a pretest leak check or for a leak check following a component change—equal to 0.00057 m³/min (0.02 cfm) or 4% of the average sampling rate, whichever is less
- θ = total sampling time, min

(b) **Case II**. One or more component changes made during the sampling run. In this case, replace V_m with the expression:

$$\left[V_m - (L_1 - L_a)\theta_1 - \sum_{i=1}^n (L_i - L_a)\theta_i - (L_p - L_a)\theta_p \right]$$

where:

 θ_i = sampling time interval, from the beginning of a run until the first component change, min

and substitute only for those leakage rates $(L_i \text{ or } L_p)$ that exceed L_a .

14.2 *Volume of Water Vapor*—Calculate the volume of water vapor of the stack gas using Equation 2.

$$V_{w(std)} = \frac{W_{lc} R T_{std}}{M_w P_{std}} = K_2 W_{lc}$$
 [Eq. 2]

where:

14.3 Volume of Moisture—Calculate the moisture content, B_{ws} , of the stack gas using Equation 3.

$$B_{ws} = \frac{V_{w(std)}}{V_{m(std)} + V_{w(std)}}$$
[Eq. 3]

where:

 B_{ws} = water vapor in the gas stream, proportion by volume

15. Calculations for Particle-Bound, Oxidized, Elemental, and Total Mercury Concentrations:

15.1 Particle-Bound Mercury

15.1.1 *Case 1: Amount of Ash on the Filter is Greater Than 0.5 g*—Calculate the concentration of mercury in $\mu g/g$ in the ash sample (Hg_{ash}) using Equation 4:

$$Hg_{ash}, \mu g/g = (IR)(DF)$$
 [Eq. 4]

where:

IR = instrument reading, $\mu g/L$

DF = dilution factor = (total digested volume, L)/(mass of ash digested, g)

Calculate the amount of mercury in the probe rinse (Hg_{pr}, Container 2) in μ g using Equation 5:

$$Hg_{nr}, \mu g = (IR)(V_1)$$
 [Eq. 5]

where:

IR = instrument reading, μ g/L

 V_1 = total volume of probe rinse sample from which sample aliquot was taken, L

Equation 5 assumes no preparation steps are needed prior to analyzing the probe rinse for mercury using CVAA. Although not required, a persulfate digest can be done on the probe rinse sample as discussed in section 13.3.3. If the persulfate digest is done equation 5 becomes Hg_{pr} , $\mu g = (IR)(V_1)(DF)$ where DF is the same as equation 9.

There is no filter blank subtraction when >0.5 g of ash are collected on the sample filter or thimble.

The total amount of particle-bound mercury (Hg_{tb}) is then determined using Equation 6:

Hg (particle),
$$\mu g = (Hg_{ash})(W_{ash}) + Hg_{pr}$$
 [Eq. 6]

where:

 W_{ash} = the total mass of ash on filter, g

The concentration of particle-bound mercury ($\mu g/dscm$) in the gas stream is then determined using Equation 7:

$$Hg^{tp}, \mu g/dscm = Hg \text{ (particle)/V}_{m(std)}$$
[Eq. 7]

where:

 $V_{m(std)}$ = is the total volume of dry gas sampled at standard (normal) conditions, dscm

15.1.2 *Case 2: Amount of Ash on the Filter is Less Than 0.5 g*—The calculation is the same as in Case 1 except the entire sample (ash and filter) is digested; therefore, DF in Equation 4 is defined only by the total digested volume. In addition a filter blank is subtracted as calculated in Equation 8.

$$Hg_{fb}, \mu g = (IR)(V_2)$$
 [Eq. 8]

where:

IR = instrument reading, $\mu g/L$

 V_2 = total volume of sample filter blank digest, L

Equation 7 for Case 2 then becomes: Hg (particle), $\mu g = (Hg_{ash})(W_{ash}) - Hg_{fb} + Hg_{pr}$

15.2 Oxidized Mercury

15.2.1 *KCl Solution (Impingers 1–3)*—Calculate the concentration of mercury in μ g/L in the KCl impinger solutions using Equation 9:

$$Hg_{KCI}, \mu g/L = (IR)(DF)$$
 [Eq. 9]

where:

IR = instrument reading, $\mu g/L$ DF = dilution factor, $\underline{V}_D + V(\underline{H}_2\underline{SO}_4) + V(\underline{HNO}_3) + V(\underline{KMnO}_4) + V(\underline{K}_2\underline{S}_2\underline{O}_8) + V(\underline{NH}_2\underline{OH})$ V_D = total digested volume, 10 mL $V(\underline{H}_2\underline{SO}_4)$ = volume of added concentrated $\underline{H}_2\underline{SO}_4$, 0.5 mL

 $V(HNO_3)$ = volume of added concentrated HNO₃, 0.5 mL

 $V(KMnO_4) =$ volume of added 5%^W/_V KMnO₄, 1.5 mL

 $V(K_2S_2O_8) = volume of added 5\%^{W/_V} K_2S_2O_4, 0.75 mL$

 $V(NH_2OH) =$ volume of added 10%^W/_V hydroxylamine sulfate, 1.0 mL

The concentration of mercury in the KCl solution blank is calculated in the same way.

15.2.2 *Total Oxidized Mercury* (Hg_0) —is defined by method as the mercury measured in the KCl sample minus the mercury measured in the KCl solution blanks, as shown in Equation 10:

$$Hg_{O}, \mu g = (Hg_{KCI})(V_{3}) - (Hg_{Ob})(V_{4})$$
[Eq. 10]

where:

 $Hg_{KCI} =$ Mercury concentration measured in KCl aliquot, $\mu g/L$

 V_3 = Total volume of aqueous KCl from which sample aliquot was taken, L

 Hg_{Ob} = Mercury concentration measured in KCl solution blank aliquot, $\mu g/L$

 V_4 = Volume of aqueous KCl originally charged to the impingers, L

The concentration of Hg^{2+} (µg/dscm) in the gas stream is then determined using Equation 11:

$$Hg^{2+}, \mu g/dscm = Hg_{O}/V_{m(std)}$$
[Eq. 11]

where:

 $V_{\mbox{\scriptsize m(std)}}$ is the total volume of dry gas sampled at standard conditions, dscm

15.3 Elemental Mercury

15.3.1 $HNO_3-H_2O_2$ Solution (Impinger 4)—Calculate the concentration of mercury in μ g/L in the HNO₃-H₂O₂ impinger solution using Equation 12:

$$Hg_{H2O2}$$
, $\mu g/L = (IR)(DF)$ [Eq. 12]

where:

IR	=	instrument reading, μg/L
DF	=	dilution factor, $\underline{V}_{D} + V(HCl) + V(\underline{H}_{2}SO_{4}) + V(KMnO_{4}) + V(\underline{K}_{2}S_{2}O_{8}) + V(NH_{2}OH)$
		V _D
V _D	=	total digested volume, 5 mL
V(HCl)	=	volume of added concentrated HCl, 0.25 mL
$V(H_2SO_4)$	=	volume of added concentrated H_2SO_4 , 0.5 mL
$V(KMnO_4)$) =	volume of added saturated KMnO ₄ , mL (volume need to turn sample to a purple
		color)
$V(K_2S_2O_8)$	=	volume of added 5% $^{W}/_{V}$ K ₂ S ₂ O ₄ , 0.75 mL (if used)
V(NH ₂ OH	=	volume of added $10\%^{W/_{V}}$ hydroxylamine sulfate, 1.0 mL

The concentration of mercury in the $HNO_3-H_2O_2$ solution blank is calculated in the same way.

15.3.2 H_2SO_4 -KMnO₄ Solution (Impingers 5-7)—Calculate the concentration of mercury in µg/L in the H_2SO_4 -KMnO₄ impinger solutions using Equation 13:

Mercury,
$$\mu g/L = (IR)(DF)$$
 [Eq. 13]

where:

DF = dilution factor, $V_D + V(HNO_3) + V(K_2S_2O_8) + V(NH_2OH)$ V_D IR = instrument reading, $\mu g/L$ V_D = total digested volume, 5 mL

 $V(HNO_3)$ = volume of added concentrated HNO₃, 0.5 mL

 $V(K_2S_2O_8) =$ volume of added 5%^W/_V K₂S₂O₄, 0.75 mL

The concentration of mercury in the H₂SO₄-KMnO₄ solution blank is calculated in the same way.

15.3.3 *Total Elemental Mercury* (Hg_E) —is defined by method as the mercury measured in the H₂SO₄–KMnO₄ impingers plus the mercury in the HNO₃–H₂O₂ impingers minus the solution blanks as shown in Equation 14:

$$Hg_{E}, \mu g = (Hg_{H2O2})(V_{4}) - (Hg_{Eb1})(V_{5}) + (Hg_{KMnO4})(V_{6}) - (H_{Eb2})(V_{7})$$
[Eq. 14]

where:

The concentration of Hg^{2+} (µg/dscm) in the gas stream is then determined using Equation 15:

$$Hg^{0}, \mu g/dscm = Hg_{E}/V_{m(std)}$$
[Eq. 15]

where:

 $V_{\text{m(std)}}$ is the total volume of dry gas sampled at standard conditions, dscm

15.4 *Total Mercury*—Is defined by the method as the sum of the particulate bound mercury, oxidized mercury, and elemental mercury as shown in Equation 16:

Hg(total),
$$\mu$$
g/dscm = Hg^{tp} + Hg²⁺ + Hg⁰ [Eq. 16]

16. Precision and Bias

16.1 Precision

16.1.1 Formal evaluation of the Ontario Hydro method was completed with dynamic spiking of Hg^0 and $HgCl_2$ into a flue gas stream.⁹ The results are shown in Table 1. The relative standard deviation for gaseous elemental mercury and oxidized mercury was found to be less than 11% for mercury concentrations greater than 3 µg/Nm³ and less than 34% for mercury concentrations less than 3 µg/Nm³. In all cases, the laboratory bias for these tests based on a calculated correction factor was not statistically significant. These values were within the acceptable range, based on the criteria established in EPA Method 301 (% RSD less than 50%).

⁹ EPRI. "Evaluation of Flue Gas Mercury Speciation Methods," EPRI TR-108988, Electric Power Research Institute, Palo Alto, CA, Dec. 1997.

16.1.2 The precision of particle-bound, oxidized, and elemental mercury sampling method data is influenced by many factors: flue gas concentration, source, procedural, and equipment variables. Strict adherence to the method is necessary to reduce the effect of these variables. Failure to assure a leak-free system, failure to accurately calibrate all indicated system components, failure to select a proper sampling location, failure to thoroughly clean all glassware, and failure to follow prescribed sample recovery, preparation, and analysis procedures can seriously affect the precision of the results.

16.2 Bias

16.2.1 Gaseous mercury species in flue gases that are capable of interacting with fly ash particles collected in the front half of the sampling train can produce a positive particle-bound mercury bias.

16.2.2 Particle-bound mercury existing in the flue gas may vaporize after collection in the front half of the sampling train because of continued exposure to the flue gas sample stream and reduced pressures during the sampling period. Such vaporization would result in a negative particle-bound mercury bias.

DRAFT

Table I

	Total V M	apor-P ercury	hase	Oxidize	ed Mer	cury	Elemental Mercury			
Ontario Hydro Method**	Mean, µg/Nm ³	Std. Dev.	RSD, %	Mean, µg/Nm ³	Std. Dev.	RSD, %	Mean, µg/Nm ³	Std. Dev.	,	
Baseline	23.35	2.05	8.79	21.24	2.13	10.02	2.11	0.65	30.69	
Hg ⁰ Spike (15.0 μg/Nm ³)	38.89	2.00	5.13	23.32	2.08	8.94	15.57	1.09	6.97	
HgCl ₂ Spike (19.9 µg/Nm ³)	42.88	2.67	6.23	40.22	2.87	7.14	2.66	0.89	33.31	

Results from Formal EPA Method 301 Evaluation Tests for the Ontario Hydro Method*

*

* For each mean result, there were 12 replicate samples (four quadtrains)
** The correction factor in all cases was not statically significant and is not shown.

17. Keywords—Air toxics, mercury, sampling, speciation

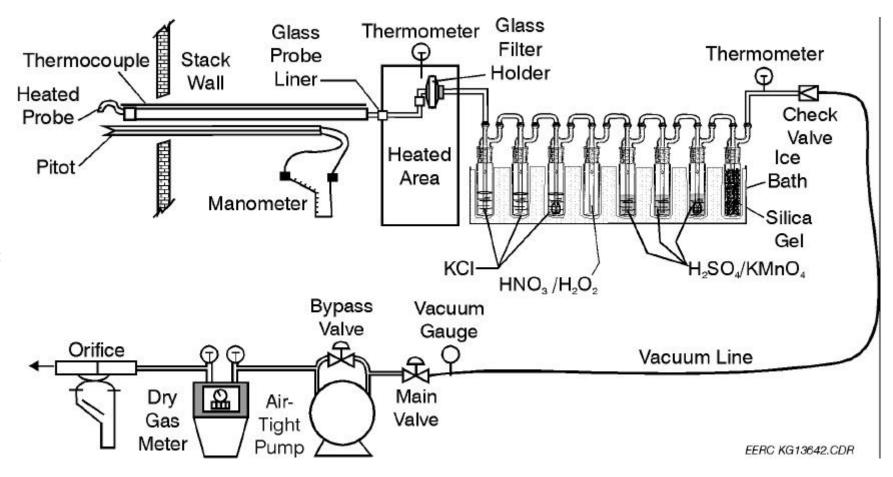


FIG. 1. Schematic of Mercury-Sampling Train in the Method 5 Configuration

Plant	Ambient Temperature °C (°F)
Location	Barometric Pressure kPa (in. Hg)
Operator	Assumed Moisture, %
Date	Probe Length, m (ft)
Run No	Nozzle Identification No.
Sample Box No	Average Calibrated Nozzle Diameter, cm (in.)
Meter Box No	Probe Heater Setting °C (°F)
Meter AH @ (kPa)	Leak Rate, m ³ /min (cfm)
C factor	Static Pressure, kPa (in. Hg)
Pitot tube coefficient, C _p	Filter No

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Schematic of Stack Cross Section

Traverse Point Number	Sampling Time	Vacuum	Stack Temp.	Velocity Head	Pressure Differential	Gas Sample Volume	Gas Sample Temp. at Dry Gas Meter		Filter Exit Temp.	Probe Exit Temp.	Final Impinger Exit Temp.
	min	kPa (in. Hg)	(T _s), °C (°F)	(△P _s) kPa (in. H ₂ O)	kPa (in. H ₂ O)	m ³ (ft ³)	Inlet °C (°F)	Outlet °C (°F)	°C (°F)	°C (°F)	°C (°F)
					L						
Total											
Average											

FIG. 2. Mercury-Sampling Field Data Report

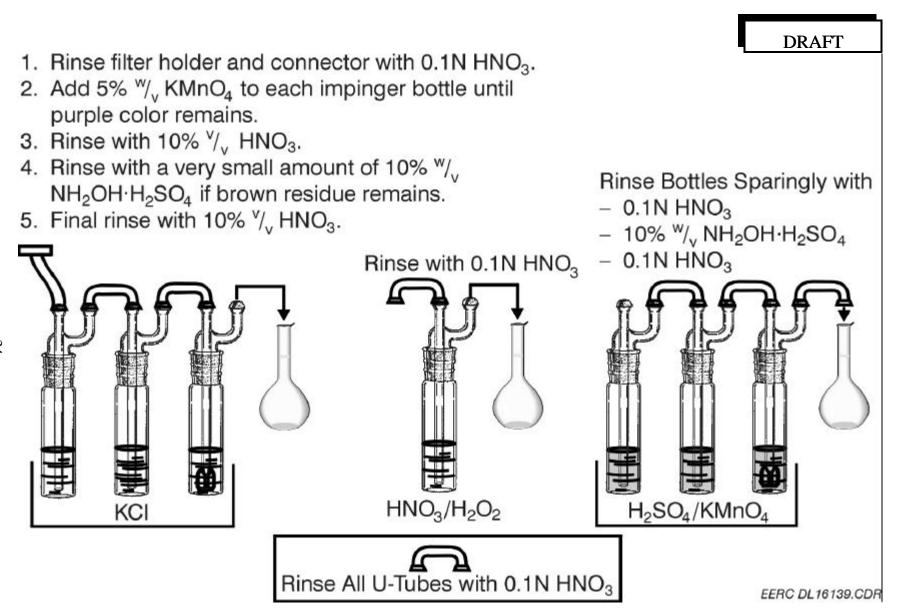


FIG. 3. Sample Recovery Scheme for the Mercury-Impinger Train

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BIBLIOGRAPHY OF EPA METHODS REFERENCED

- (1) Method 303F in *Standard Methods for the Examination of Water and Wastewater*, 15th Edition, 1980. Available from the American Public Health Association, 1015 18th Street N.W., Washington, D.C. 20036.
- (2) EPA Methods 6010, 6020, 7000, 7041, 7060, 7131, 7421, 7470, 7740, and 7841, *Test Methods for Evaluating Solid Waste: Physical/Chemical Methods.* SW-846, Third Edition. September 1988. Office of Solid Waste and Emergency Response, U. S. Environmental Protection Agency, Washington, D.C. 20460.
- (3) EPA Methods 1 through 5, *Code of Federal Regulations*, Title 40, Part 60, Appendix A, July 1, 1991.
- (4) EPA Method 101A, *Code of Federal Regulations*, Title 40, Part 61, Appendix B, July 1, 1991.
- (5) EPA Method 29, *Emission Measurement Technical Information Center*, EMTIC TM-029, April 25, 1996.
- (6) U.S. Environmental Protection Agency "Method 301 Field Validation of Pollutant Measurement Method from Various Waste Media," In *Code of Federal Regulations*, Title 40, Parts 61 to 80. Washington, DC, USA, Office of the Federal Register, Part 63, Appendix A, pp 324–331, July 1993.